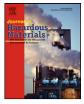
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Journal of Hazardous Materials



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Chromium recovery from exhausted baths generated in plating processes and its reuse in the tanning industry

Josep Torras^a, Irene Buj^{b,*}, Miquel Rovira^{c,d}, Joan de Pablo^{c,d}

^a Department of Mining Engineering and Natural Resources, Universitat Politècnica de Catalunya, Av. Bases de Manresa, 61-73, 08240, Manresa, Barcelona, Spain

^b Department of Mechanical Engineering, Universitat Politècnica de Catalunya, Barcelona, Spain

^c CTM Technological Centre, Universitat Politècnica de Catalunya, Av. Diagonal, 647, 08028, Manresa, Barcelona, Spain

^d Department of Chemical Engineering, Universitat Politècnica de Catalunya, Av. Diagonal, 647, 08028, Barcelona, Spain

ARTICLE INFO

Article history: Received 18 July 2011 Received in revised form 14 November 2011 Accepted 10 January 2012 Available online 21 January 2012

Keywords: Chromium recovery Exhausted baths Surface treatment Tanning Chromium reduction

ABSTRACT

Chromium plating used for functional purposes provides an extremely hard, wear and corrosion resistant layer by means of electrolytic deposition. Typical layer thicknesses range between 2.5 and 500 μ m. Chromium electroplating baths contain high concentrations of Cr(VI) with chromium trioxide (CrO₃) as the chromium source. When because of technical or economic reasons a bath gets exhausted, a waste containing mainly chromium as dichromate as well as other heavy metals is generated. Chromium may then be purified for use in other industrial processes with different requirements. In this work, a sustainable system for using galvanic wastes as reagents in the leather tanning industry, thus reducing quantity of wastes to be treated, is presented. Metal cations present in the chromium (III) sulphate. From chromium sulphite in acidic conditions, Cr(VI) was reduced to Cr(III) as chromium (III) sulphate. From chromium (III) sulphate may be obtained, which is one of most used compounds in the taning industry. Cr(III) concentration in the final solution allows its reuse without concentration, but with a slight dilution.

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1. Introduction

Hard or functional chromium plating of metallic parts provides an extremely hard and corrosion resistant layer with low friction coefficient and exceptional wear resistance. It may also be used for rebuilding worn parts like rolls, dies or tools. Usual thicknesses of the chromium layer ranges between 2.5 and 500 μ m [1].

Electrolytic chromium plating baths have a high concentration of Cr(VI) as dichromate owing to the high acidity of the system because of chromic acid and sulphuric acid. Usual chromium concentration ranges between 110 and 130 g/l of chromium, which is equivalent to about 150 g/l Cr₂O₃ [1]. During bath operation it is necessary to increase chromium concentration so that the bath does not lose effectiveness. Nevertheless, with time impurities accumulate in the bath due to electrochemical reactions as well as oil and particles on the workpieces' surface. Finally, after a long time the bath gets exhausted and needs to be changed. The problem with rinsing waters is even more important than that of exhausted baths. Great volumes of water are generated containing chromium at such a low concentration that it is difficult to completely reuse it. The tanning process transforms skins into leather by incorporation of chromium in its structure. Basic chromium sulphate, $CrSO_4(OH)$ is a common salt employed in the tanning industry. In order to obtain it, chromium dichromate, $K_2Cr_2O_7$, is used. Hexavalent chromium is reduced to trivalent chromium in acidic conditions. Later, in order to increase the tanning properties of chromium salts and to improve its fixation on leather, it is usual to basify them. To do so, anhydrous sodium carbonate is added to the solution. Usual initial chromium concentration is 20 g/l[2]. Typical spent tanning baths Cr(III) concentrations range between 3.5 and 4 g/l, with neutral salts such as NaCl between 50 and 60 g/l and Na₂SO₄ between 10 and 12 g/l [3].

For chromium reduction in origin in the surface treatment industry, for instance cleaning process of racks, use of pure metal anodes and filtration of the bath solution are recommended [4]. For chromium recovery, hydrometallurgical techniques such as liquid–liquid extraction, ion exchange and/or electrochemical separation [5] or precipitation, acid leaching and subsequent precipitation [6,7] are recommended.

In the tanning industry, although some authors describe application of biosurfactants [8] and also recycling of spent tanning liquors after membrane separation, usual recovery method consists of Cr(III) is precipitation as chromic hydroxide by means of an alkali. Later, Cr(III) in the concentrated sludge is dissolved with sulphuric acid and basic chromic sulphate is regenerated for reuse.

^{*} Corresponding author. Tel.: +34 93 4054016; fax: +34 93 4016693. *E-mail address:* irene.buj@upc.edu (I. Buj).

^{0304-3894/\$ -} see front matter © 2012 Elsevier B.V. All rights reserved. doi:10.1016/j.jhazmat.2012.01.036

However, the method does not solve the problem of dissolved neutral salts in the liquor [9]. In addition, evaporation of the solution is necessary in this case so as to achieve desired Cr(III) concentration of 20 g/l.

Other authors obtained basic chromium sulphate from wastes from the tanning industry different than exhausted tanning liquors. Erdem studied the reduction of chromate ions from chrome shaving generated in tanning processes, by adjustment of pH with H_2SO_4 . Contact time was 30 min and at the end of the process a solution of chromium sulphate was obtained [10]. On the other hand, Dettmer et al. employed H_2SO_4 to obtain $Na_2Cr_2O_7$ from Na_2CrO_4 from heat ashes of thermally treated leather. Later, they reduced dichromate with Na_2SO_3 and obtained chromium (III) sulphate, $Cr_2(SO_4)_3$. From that compound, by adjustment of pH with sodium hydroxide or sodium bicarbonate, they obtained basic chromium sulphate [11].

Basic chromium sulphate could alternatively be obtained from galvanic exhausted baths containing Cr(VI), provided that after the reduction process chromium is found as an appropriate compound and at the right concentration. Thus, tanning industry can reuse chromium that the surface treatment industry is not able to use. Many processes employed for reducing Cr(VI) concentrations in solution are based on reduction of Cr(VI) to Cr(III) and precipitation of Cr(III) [12,13]. Different reducing agents are known for chromium. They are classified into four main groups: sulphur compounds, Fe⁰ or Fe²⁺ compounds, organic compounds and other. Amongst sulphur compounds, hydrogen sulphide (H₂S) [14], iron (II) sulphide (FeS₂) [15], sodium metabisulphite (Na₂S₂O₅) [16] and sodium sulphite (Na₂SO₃) have been reported. Regarding iron compounds, metallic Fe [17,18], ammonium iron (II) sulphate $(Fe(NH_4)_2(SO_4)_2)$ [19], sodium hydrosulphide (NaHS) and iron (II) sulphate (FeSO₄) [20] and siderite (FeCO₃) [21] amongst other are known. Concerning organic compounds, common ones are ethylene glycol [22], ascorbate [23], citrate [24], organic acids with TiO₂ [25] or soil organic matter [26]. Other reagents such as H₂O₂ have been also employed [8].

The present work shows a sustainable method for recovering chromium from exhausted chromium plating baths generated by the surface treatment industry and their further use in the tanning industry. Exhausted baths show a high chromium concentration (about 110 g/l Cr) [1]. After appropriate treatments, it will lead to a Cr(III) solution with a higher concentration than that of tanning liquors (about 20 g/l Cr) [2]. Thus, an evaporation process to concentrate the solution is not necessary in this case.

Cost of basic chromium (III) sulphate is relatively low (about $1 \in$ per kg of powder material) [27]. However, the strategy of reusing wastes from a certain industry as raw material for another productive process falls squarely within the field of industrial ecology [28]. By means of such strategy quantities of hazardous wastes generated in the surface treatment industry will be reduced. That is the recommended option by the United Nations Environment Programme's (UNEP's) International Panel for Sustainable Resource Management (Resource Panel) on metal recycling rates [29]. Although the present method requires a reduction reaction of Cr(VI) to Cr(III), such reduction would also be necessary if exhausted chromium plating baths were treated by stabilization with Portland cement [30], since chromium is in anionic form.

2. Materials and methods

2.1. Materials

A sample from an exhausted bath from hard chromium plating was provided by a surface treatment company in the area of Barcelona. Reagents used in this work, NaOH, Na₂SO₃ and

Table 1

Molarity, mass (g) and volume (ml) of the different compounds employed in the reduction reaction: a) stoichiometric quantities; b) excess of Na₂SO₃.

	Molarity	Mass (g)	Volume (ml)
a) Stoichiometric quantities			
Cr neutralized bath	0.065		30.000
H_2SO_4	0.163	15.974	8.870
Na ₂ SO ₃	0.098	12.348	
b) Excess of Na ₂ SO ₃			
Cr neutralized bath	0.065		30.000
H_2SO_4	0.170	16.660	9.260
Na ₂ SO ₃	0.170	21.420	

H₂SO₄ (95–97%), were purchased from Scharlau (analytical grade). Deionized water was obtained by means of Milli-Q system. Water resistivity was 18.2 M Ω cm at 25 °C. It was filtered with a 0.22 μ m pore diameter filter. Ultra-trace quality nitric acid was used for preparing standard solutions. ICP-MS (inductively coupled plasma mass spectrometer) calibration standards for quantitative analysis of Cr, Cd, Cu, Ni, Pb and Zn, as well as for semi quantitative analysis of Mg, Rh and Pb were provided by Scharlau. Metal concentration of all standards was 1000 mg/l.

2.2. Experimental procedure

A process for recovery of chromium from surface treatment exhausted baths and subsequent reuse in the tanning industry was developed. The process consists of following steps: bath characterization, bath neutralization with NaOH and Cr(VI) reduction to Cr(III) with Na₂SO₃ in acidic conditions.

After characterization of the bath, neutralization was performed. To that end, 7.027 g of NaOH was added to 50 ml of the exhausted bath. A precipitate containing metal hydroxides was formed. The neutralized bath was filtered in order to remove the precipitate.

For the reduction reaction, 30 ml of the neutralized bath containing Na_2CrO_4 were mixed with Na_2SO_3 in solution. The reduction reaction of sodium chromate in solution is as follows (Eq. (1)):

$$2Na_2CrO_4 + 3Na_2SO_3 + 5H_2SO_4 \rightarrow Cr_2(SO_4)_3 + 5Na_2SO_4 + 5H_2O(1)$$

Two different conditions were considered: stoichiometric quantities (molar ratio of 3 to 2 with respect to Na_2CrO_4 according to Eq. (1)) and excess of Na_2SO_3 (molar ratio of 5.2 to 2 with respect to Na_2CrO_4) (Table 1).

Reaction was considered to be finished after 30 min.

2.3. Methods of analysis

The exhausted bath was characterized by means of elemental semiquantitative analysis, elemental quantitative analysis of Cd, Cr, Cu, Ni, Pb and Zn, pH measurement and titration. Elemental analyses were performed in a Perkin Elmer ELAN 6000 ICP-MS. Samples were diluted with HNO₃ (Scharlau, trace analysis) at 1%. pH was measured by means of a CRISON GLP 22 pH-meter with temperature compensation. Titration was performed with NaOH 2 N.

Both the neutralized bath and the final solution were characterized by means of elemental quantitative analysis and pH measurement.

Hexavalent chromium in the final solution was determined by diphenylcarbazide method [31] using a UV–vis spectrometer (Shimadzu model UV-1603).

Table 2
Cr, Ni, Cd, Pb, Cu and Zn concentration (mg/l) and pH value in the exhausted bath.

	Exhausted bath	
[Cr(VI)] (mg/l)	113079.0	
[Ni] (mg/l)	85.2	
[Cd] (mg/l)	5.0	
[Pb] (mg/l)	51.6	
[Cu] (mg/l)	2449.0	
[Zn] (mg/l)	1633.0	
pH	0.87	

2.4. Geochemical models

In order to determine the solubility behaviour of chromium in both the neutralized bath and the final solution used for tanning, the system was modelled with the geochemical code Hydra/Medusa [32]. It is a chemical equilibrium model for calculation of metal speciation. pH–fraction diagrams were obtained for both the exhausted bath and the final solution.

3. Results

3.1. Bath characterization

In Table 2 Cd, Cr, Cu, Ni, Pb and Zn concentration in the exhausted bath, as well as pH, are shown.

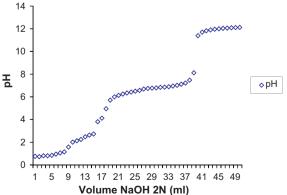
pH of the bath is extremely acid. It contains mainly hexavalent chromium and also divalent metals, such as Cu and Zn, at a lower concentration. The presence of other metals at significant concentrations was discarded by means of elemental semiquantitative analysis.

For the titration, 50 ml of the exhausted bath was employed. NaOH 2 N was added to the bath in order to determine the buffering pH (Fig. 1).

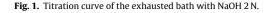
In the central part of the curve, where a volume between 20 and 40 ml NaOH 2 N was added, there is a plateau. It corresponds to neutral pH, between 6 and 8, where most divalent cations precipitate as hydroxides.

Fraction diagram in Fig. 2, obtained with the geochemical code Hydra/Medusa, shows species distribution of the exhausted bath at different pH values.

When pH raises from 6 to 8 dichromates and hydrogen chromates are transformed into chromates. Once transformation has finished, more NaOH will lead to an increase of pH.



Titration curve of the exhausted bath



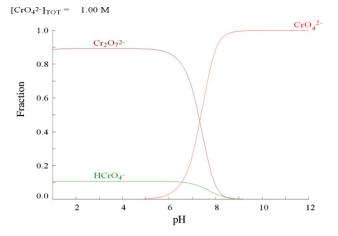


Fig. 2. pH-fraction diagram of the chromium plating exhausted bath.

Table 3

Cr, Ni, Cd, Pb, Cu and Zn concentration (mg/l) and pH value in the neutralized exhausted bath.

	Neutralized solution
[Cr(VI)] (mg/l)	112423.0
[Ni] (mg/l)	3.3
[Cd] (mg/l)	0.4
[Pb] (mg/l)	<0.05
[Cu] (mg/l)	52.2
[Zn] (mg/l)	44.1
рН	6.64

3.2. Bath neutralization

Taking into account the previously obtained titration curve, the necessary quantity of NaOH was added to the bath to obtain neutral pH (plateau in the central part of the curve). This allows precipitation of most heavy metals as hydroxides, avoiding resolubilization due to amphoteric behaviour of metal hydroxides. Moreover, extremely basic pH is not desirable, since the subsequent reduction reaction needs acidic conditions.

In Table 3 metal concentrations and pH of the neutralized bath are presented.

The level of precipitation for divalent metals is calculated as follows (Eq. (2)).

$$LP(\%) = \frac{C_0 - C_f}{C_0} \times 100$$
(2)

where LP is the level of precipitation (%), C_0 is the initial concentration of a metal in solution (mg/l), and C_f is the final concentration of a metal in solution (mg/l).

Results in Table 4 show that level of precipitation is higher than 92% for all metals studied, with an extremely high value of 99.9% for Pb. After bath neutralization with NaOH, divalent metal concentrations were reduced as expected. Nevertheless, neutralization did not reduce Cr concentration markedly, since Cr is found in the anion form.

Table 4 LP (%) of divalent metals after neutralization with NaOH.

Divalent metal	LP (%)	
Ni	96.1	
Cd	92.0	
Pb	99.9	
Cu	97.9	
Zn	97.3	

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Table 5

Cr(VI) concentration (mg/l) in the final solution.

Reaction	Cr(VI) concentration (mg/l)
Stoichiometric quantities	2902.000
Excess Na ₂ SO ₃	<0.003

Table 6

Cr, Ni, Cd, Pb, Cu and Zn concentration (mg/l) and pH value in the final solution.

	Final solution	
[Cr(III)] (mg/l)	32250.0	
[Ni] (mg/l)	<1.0	
[Cd] (mg/l)	0.3	
[Pb] (mg/l)	<0.05	
[Cu] (mg/l)	4.7	
[Zn] (mg/l)	11.3	
рН	2.33	

3.3. Final solution

In order to assure that all Cr(VI) had been reduced to Cr(III), Cr(VI) concentration was determined, for both stoichiometric quantities of Na_2SO_3 (molar ratio of 3 to 2 with respect to Na_2CrO_4 according to Eq. (1)) and excess Na_2SO_3 (molar ratio of 5.2 to 2 with respect to Na_2CrO_4) (Table 5).

No hehavalent chromium was detected by UV–vis after the redox reaction with excess of Na_2SO_3 , showing the excellent efficiency of the reduction process. According to results, excess Na_2SO_3 is recommended in the reduction reaction.

In Table 6 heavy metals concentration and pH are presented for the final solution after reduction reaction.

Taking into account that after the reduction reaction there was a dilution of 28.7%, divalent metal concentrations remained almost constant.

In order to determine chemical species in solution, a pH–fraction diagram was obtained from Hydra-Medusa geochemical software [32] (Fig. 3). In the X axis pH is presented, whilst in the Y axis fraction of the different species is shown. Cr(III) concentration in the final solution was 32.25 g/l (Table 6), which corresponds to 0.62 mol/l of CrO_4^{2-} , and sodium sulphite concentration was 1.61 mol/l (excess Na_2SO_3).

Cr(OH)SO₄ is found around pH 3. The chromium sulphate solution may be basified with sodium hydrogen carbonate in order to use it for leather tanning [10]. On the other hand, final Cr concentration of 32250 mg/l was even higher than that requested by the tanning industry (about 20,000 mg/l) [2]. Therefore, a slight dilution would be necessary before using the final solution for tanning.

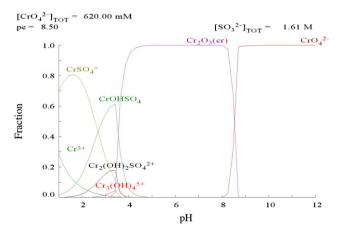


Fig. 3. pH-fraction diagram of the final solution.

4. Conclusions

In this work a sustainable process to obtain a tanning solution from an exhausted plating bath is presented. The bath was first neutralized with sodium hydroxide for the precipitation of divalent metals as hydroxides. Then, Cr(VI) in the neutralized bath was reduced to Cr(III) with an excess of sodium sulphite. The resulting chromium sulphate solution can be basified with sodium hydrogen carbonate prior to its reuse in the leather tanning industry.

The process allows reduction of quantities of hazardous wastes generated by the galvanic industries by means of its reuse in the tanning industry. That is the preferred option by United Nations Environment Programme's International Panel for Sustainable Resource Management (Resource Panel) on metal recycling rates. Unlike direct reuse of exhausted tanning liquors, reuse of exhausted galvanic baths reduces impurities in the final bath. Moreover, final chromium solution is high enough to avoid a concentration operation in an evaporator.

Acknowledgement

The authors thank the Spanish Ministry of Science and Technology (REN2002-02971/TECNO) for financial help.

References

- [1] K.R. Newby, Functional chromium plating, Met. Finish. 98 (2000) 223-233.
- [2] H. Tahir, Comparative trace metal contents in sediments and liquid wastes from tanneries and the removal of chromium using zeolite-5A, Electron. J. Environ. Agric. Food Chem. 4 (4) (2005) 1021–1032.
- [3] D. Petruzzelli, R. Passino, G. Tiravanti, Ion exchange process for chromium removal and recovery from tannery wastes, Ind. Eng. Chem. Res. 34 (1995) 2612–2617.
- [4] J.R. Viguri, A. Andrés, A. Irabien, Waste minimization in a hard chromium plating small medium enterprise (SME), Waste Manage. 22 (2002) 931–936.
- [5] A.G. Chmielewski, T.S. Urbanski, W. Migdal, Separation technologies for metals recovery from industrial wastes, Hydrometallurgy 45 (1997) 333–344.
- [6] B. Gupta, A. Deep, S.N. Tandon, Recovery of chromium and nickel from industrial waste, Ind. Eng. Chem. Res. 41 (2002) 2948–2952.
- [7] P.T. de Souza e Silva, N.T. de Mello, M.M. Menezes Duarte, M.C.B.S.M. Montenegro, A.N. Araújo, B. de Barros Neto, et al., Extraction and recovery of chromium from electroplating sludge, J. Hazard. Mater. 128 (2006) 39–43.
- [8] E. Kiliç, J. Font, R. Puig, S. Çolak, D. Çelik, Chromium recovery from tannery sludge with saponin and oxidative remediation, J. Hazard. Mater. 185 (2011) 456–462.
- [9] K.J. Sreeram, T. Ramasami, Sustaining tanning process through conservation, recovery and better utilization of chromium, Resour. Conserv. Recycl. 38 (2003) 185–212.
- [10] M. Erdem, Chromium recovery from chrome shaving generated in tanning processes, J. Hazard. Mater. B129 (2006) 143–146.
- [11] A. Dettmer, K.G.P. Nunes, M. Gutterres, N.R. Marcílio, Production of basic chromium sulfate by using recovered chromium from ashes of thermally treated leather, J. Hazard. Mater. 176 (2010) 710–714.
- [12] F. Demoisson, M. Mullet, B. Humbert, Pyrite oxidation by hexavalent chromium: investigation of the chemical processes by monitoring of aqueous metal species, Environ. Sci. Technol. 39 (2005) 8747–8752.
- [13] A.K. Golder, A.K. Chanda, A.N. Samanta, S. Ray, Removal of hexavalent chromium by electrochemical reduction-precipitation: investigation of process performance and reaction stoichiometry, Sep. Purif. Technol. 76 (2011) 345–350.
- [14] B. Hua, B. Deng, Influences of water vapor on Cr(VI) reduction by gaseous hydrogen sulfide, Environ. Sci. Technol. 37 (2003) 4771–4777.
- [15] Y. Lin, C. Huang, Reduction of chromium(VI) by pyrite in dilute aqueous solutions, Sep. Purif. Technol. 63 (2008) 191–199.
- [16] J.P. Beukes, J. Pienaar, G. Lachmann, E.W. Giesekke, The reduction of hexavalent chromium by sulphite in wastewater, Water S.A. 25 (1999) 363–370.
- [17] A. Fiúza, A. Silva, G. Carvalho, A.V. de la Fuente, C. Delerue-Matos, Heterogeneous kinetics of the reduction of chromium (VI) by elemental iron, J. Hazard. Mater. 175 (2010) 1042–1047.
- [18] P. Mitra, D. Sarkar, S. Chakrabarti, B.K. Dutta, Reduction of hexa-valent chromium with zero-valent iron: batch kinetic studies and rate model, Chem. Eng. J. 171 (2011) 54–60.
- [19] L.E. Eary, D. Rai, Chromate removal from aqueous wastes by reduction with ferrous ion, Environ. Sci. Technol. 22 (1988) 972–977.
- [20] G. Qin, M.J. McGuire, N.C. Blute, C. Seidel, L. Fong, Hexavalent chromium removal by reduction with ferrous sulfate, coagulation, and filtration: a pilotscale study, Environ. Sci. Technol. 39 (2005) 6321–6327.

- [21] M. Erdem, F. Gur, F. Tumen, Cr(VI) reduction in aqueous solutions by siderite, J. Hazard. Mater. 113 (2004) 217–222.
- [22] P.R. Bontchev, A. Malinovski, M. Mitewa, K. Kabassonov, Intermediate chromium(V) complex species and their role in the process of chromium(VI) reduction by ethylene glycol, Inorg. Chim. Acta 6 (1972) 499–503.
- [23] P. O'Brien, N. Woodbridge, A study of the kinetics of the reduction of chromate by ascorbate under aerobic and anaerobic conditions, Polyhedron 16 (1997) 2081–2086.
- [24] C. Li, Y. Lan, B. Deng, Catalysis of manganese(II) on chromium(VI) reduction by citrate, Pedosphere 17 (2007) 318–323.
- [25] B. Deng, A.T. Stone, Surface-catalyzed chromium (VI) reduction: The TiO₂-CrVI-mandelic acid system, Environ. Sci. Technol. 30 (1996) 463–472.
- [26] M.K. Banks, A.P. Schwab, C. Henderson, Leaching and reduction of chromium in soil as affected by soil organic content and plants, Chemosphere 62 (2006) 255–264.

- [27] http://www.alibaba.com/Chemicals_p8, 2011.
- [28] R.A. Frosch, N. Gallopoulos, Strategies for manufacturing, Sci. Am. 261 (3) (1989) 144–152.
- [29] T.E. Graedel, J. Allwood, J.P. Birat, M. Buchert, C. Hagelüken, B.K. Reck, S.F. Sibley, et al., What do we know about metal recycling rates? J. Ind. Ecol. 15 (3) (2011) 355–366.
- [30] J.R. Conner, Chemical Fixation and Solidification of Hazardous Wastes, Van Nostrand Reinhold, New York, 1990.
- [31] USEPA, Method 7196A: chromium, hexavalent (colorimetric), in: Test Methods for Evaluating Solid Waste, Physical/Chemical Methods, SW-846, 3rd ed., U.S. Environmental Protection Agency, 1992, 6.
- [32] I. Puigdomènech, HYDRA/MEDUSA. Chemical Equilibrium Software, 2004.